

# Species Distribution Modeling and Ecological Niche Modeling: Getting the Concepts Right

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## Abstract

We provide an overview of conceptual considerations in terminology related to ecological niche modeling and species distribution modeling, two near-synonymous (but not quite), relatively new tools in macroecology and biogeography. We show that a large majority of published studies taking advantage of these tools use terminology inappropriate to the biogeographic and ecological basis on which their application is founded. We suggest that only via rigorous and appropriate terminology will these tools achieve their fullest potential.

**Key words:** Geographic Range, Niche, Species Distributions, Macroecology, Biogeography.

## Introduction

Interest in describing, understanding, and predicting geographic and environmental distributions of species is very old (Wallace 1860; Grinnell 1917). In the last 20 years, methods have been developed to estimate distributional areas on the basis of correlations of known occurrences with environmental variables. Usage of these methodologies has literally exploded in recent years, now with hundreds of papers being published every year (Lobo *et al.* 2010). Both the fundamental importance of distributional areas as a concept in biogeography and ecology, and myriad applications to which these methods can be applied explain the massively increasing interest in this field (Peterson *et al.* 2011).

Since the late 1990s, two terminologies have been used to refer to correlative summaries of species' environmental associations and the relationships of those associations to their geographic distributions: "species distribution models" (SDM; Elith & Leathwick 2009; Franklin 2010) and "ecological niche models" (ENM; Harrison 1997; Peterson *et al.* 1999); other, more neutral terms have also been used, such as "bioclimatic envelope models" (Araújo & Peterson 2012). Arguments regarding these terms have been numerous (Peterson 2006; Elith & Leathwick 2009; Franklin 2010; Sillero 2011; Araújo & Peterson 2012; Warren 2012) and rather inconclusive. In general, it appears that those who use SDM prefer to avoid overinterpreting the ecological significance of the model, or worry that many dimensions of true ecological niches remain uncharacterized by these methods. The ENM world has emphasized that

their focus is on a subset of all ecological niche dimensions that is defined in coarse-resolution dimensions that are relatively unlinked to (i.e., not affected by) the population processes of the species in question (Soberón 2007; Soberón & Nakamura 2009; Soberón 2010). A first in-depth synthesis of concepts in the field was offered recently (Peterson *et al.* 2011), but has not as-yet seen broad uptake.

The debate between ENM and SDM is far from being merely semantic. It is perfectly feasible to model species' distributions without resorting to a niche definition or even referring to any environmental variables (Jennrich & Turner 1969; Rapoport 1982; Bahn & McGill 2007). On the other hand, modeling the *processes* that produce and shape the area of distribution, transferring causal factors in time or in space, or interpreting biologically the obtained *pattern*, obviously requires some hypothesis about the ecology of the species, which is clearly a niche-related inquiry.

In this paper, we discuss this dichotomy of terminology. To characterize how these terminologies are used in the field, we searched Web of Science for journal articles that used "species distribution model" or "ecological niche model" under topic or title. Of the 242 papers that resulted, 13 were by one or the other of us, and so were removed from consideration. A further three used the term SDM, but in a very different sense more closely related to environmental chemistry (Lee *et al.* 1995; Shen *et al.* 1995; Shen & Lin 2003); these papers were also removed from analysis.

Because these questions involve subtleties of how terms are used, we focused on the abstract associated with each publication, and the terminology used therein. Of the 226 papers analyzed, 127 used SDM, 51 used ENM, and 50 were

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nonspecific or vague about the concept used or used other terminologies (Table 1). Indeed, two papers used *both* terms (Colacicco-Mayhugh *et al.* 2010; Roubicek *et al.* 2010)! As will be analyzed in greater depth below, no clear divisions existed regarding the uses to which these models were put, such that – in effect – we see simply different terms being used for the same thing.

## Conceptual Framework and Real-world Implications

All of these questions regarding the complexities of ecological niches and geographic distributions require careful and consistent terminology (Peterson *et al.* 2011). Although other ‘vocabularies’ have been offered (Godsoe 2010a; Sillero 2011), they have not successfully captured the complexity of the situation: species are distributed in both geographic and environmental dimensions (Colwell & Rangel 2009), and both geographic and ecological distributions must be captured and linked in any effective terminology. We use the Biotic-Abiotic-Mobility (BAM; Figure 1) framework (Soberón & Peterson 2005) to contemplate the suites of factors that determine geographic distributions of species, and that have been known and discussed for many decades (Grinnell 1924; Good 1931). This view of species’ distributional ecology captures and frames many of the key phenomena for which models of niches and distributions have been used: characterizing niches; interpolating distributional patterns; and anticipating unknown distributional areas, geographic potential of invasive species, and responses to changing environmental conditions.

We contemplate two related spaces. In geographic space  $\mathbf{G}$ , we can consider several different distributional areas. Areas appropriate in terms of abiotic conditions can be termed  $\mathbf{G}_A$ , or the set of areas that meet the conditions of the fundamental niche  $\mathbf{N}_F$  of the species. A similar set of constraints applies to biotic conditions, wherein a subset of  $\mathbf{G}$  is identified that meets the suite of biotic conditions necessary for the maintenance of populations of the species. Areas that have been accessible to the species over relevant time periods can be termed  $\mathbf{M}$ , which has been discussed in detail in a previous contribution (Barve *et al.* 2011). Areas that fulfill both the abiotic and biotic requisites of the species constitute the potential geographic distribution of the species  $\mathbf{G}_p$ , and the intersection of the potential distribution with the accessible areas is termed the occupied

distributional area  $\mathbf{G}_O$ ; finally, the presently-inaccessible portions of the potential distributional area are termed the invadable area  $\mathbf{G}_i$ .

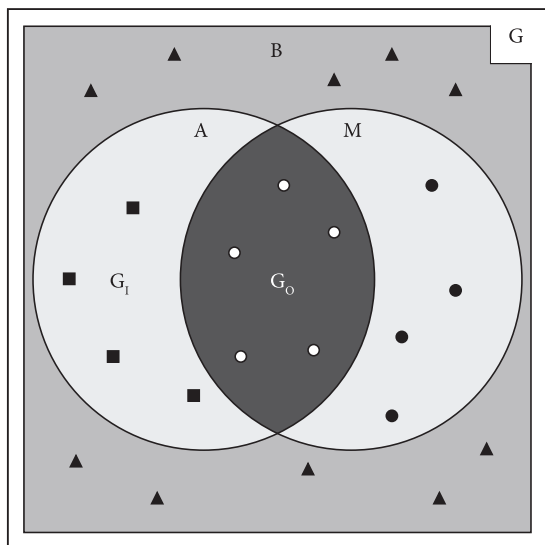
Besides  $\mathbf{G}$ , another space is required: in the multivariate space of environmental dimensions, the environments associated with  $\mathbf{G}$  are  $\eta(\mathbf{G})$ , or  $\mathbf{E}$ . Each of the distributional areas discussed above has a corresponding entity in environmental space: to refer to corresponding elements in environmental space of any distributional area, we use the notation  $\mathbf{E}' = \eta(\mathbf{G}')$ ; its converse operation is  $\eta^{-1}(\mathbf{E}')$ , which identifies the geographic location(s) that correspond to a given environmental combination  $\mathbf{E}'$ .

The fundamental niche  $\mathbf{N}_F$ , however, is not generally fully represented in real-world geography – that is, the fundamental niche is a construct of physiological responses, and may extend to sets of conditions not represented within the study area. The subset of the fundamental niche that is actually represented on relevant landscapes, or the existing fundamental niche, can be defined as  $\mathbf{N}_F \cap \eta(\mathbf{M}) = \mathbf{N}_F^*$  (Jackson & Overpeck 2000), and indeed this estimate may be reduced still further if sampling is incomplete across the landscape: if the area sampled is  $\mathbf{S}$ , then we estimate a yet-different set of environments  $\hat{\mathbf{N}}_F^* = \mathbf{N}_F \cap \eta(\mathbf{M} \cap \mathbf{S}) \subseteq \mathbf{N}_F^*$ . This reduction of the fundamental niche is critical in the clarifications offered in this paper, as it basically indicates that any ecological niche that can be estimated on real-world landscapes for real-world species will be only partial, and that these niches can be compared only as conditioned on the  $\mathbf{M}$  and  $\mathbf{S}$  areas corresponding to each of the species in question (Jiménez-Valverde *et al.* 2008).

Note that some crucial points have been made here. If ‘niche modeling’ were genuinely an attempt to estimate fundamental niches, which could then be used to outline  $\mathbf{G}_A$ , then they would *not* be estimating actual distributional areas, but rather  $\mathbf{G}_p$ , or the potential distribution of the species. Hence, niche model outputs must be processed further (i.e., reduction to the portion of  $\mathbf{G}_A$  that is within  $\mathbf{M}$ ) if an actual (occupied) distributional area is to be identified. Second, on the environmental side, these approaches at best can characterize  $\mathbf{N}_F^*$ , which is the fundamental niche conditioned on the associated  $\mathbf{M}$  and  $\mathbf{S}$  – these niche estimates are not comparable without explicit reference to  $\mathbf{M}$  and  $\mathbf{S}$ . These two points have massive implications for what researchers are ‘doing’ with these techniques at the moment.

**Table 1.** Summary of uses to which different studies have put analyses under the rubrics of ‘ecological niche modeling’ and ‘species distribution modeling’.

	Ecological niche modeling	Species distribution modeling	Total
Climate change projections	12	35	47
Distributional prediction	19	62	81
Species invasions projections	7	22	29
Niche characterization	2	3	5
Paleodistributional predictions	11	4	15
Total	51	126	177



**Figure 1.** A BAM diagram configuration that may be most relevant to broad geographic questions about species' geographic distributions. That is, as argued by Soberón (2007), biotic considerations may frequently be diffuse and non-limiting, or may be manifested at spatial resolutions so fine as to be nestled within the coarse-resolution phenomena that are the focus of most ecological niche modeling and species distribution modeling studies. The circle **A** represents the parts of the world that contain the abiotic conditions required for a species' survival and growth. The circle **M** represents the region that has been accessible to the species over a relevant period of time. The intersection of these two regions is the occupied area  $G_o$ ;  $G_i$  holds the correct suite of environmental conditions, but has not been explored by the species; open circles represent presence data; closed circles indicate absences due to incorrect environment; closed squares indicate absences due to lack of dispersal capacity; and triangles indicate absences owing to both incorrect conditions and limited dispersal. Note that absences can also occur within  $G_o$  as well. From Saupe *et al.* (2012).

## What is the Target of Modeling?

Some early thinking approaches to understanding species' distributions were developed in exclusively spatial dimensions, while others were developed in exclusively environmental dimensions (Hutchinson 1957; Brown *et al.* 1996). Indeed, space-only approaches to mapping species' distributions continue strong in the field of spatial epidemiology, in which, in many studies, disease distributions are estimated and characterized only in dimensions of space, and frequently with no reference to environmental conditions (Pfeiffer *et al.* 2008; Lai *et al.* 2009). The environment-only tradition continues also, but quietly, hidden as an unstated assumption of many studies under the rubric of ecological niche modeling or species distribution modeling (see "When are Species' Distributions Modeled Correctly?" below).

In essentially all modern applications of these techniques, however, the process involves at least two transitions between spaces. That is, known occurrences,  $G_+$ , are distributed across geographic space. The modeling process involves

comparing environments associated with occurrences  $\eta(G_+)$  with those associated with some set of data representing non-occurrences. Hence, the model is developed in environmental space, and the target of modeling is some subset of **E**, and not an area of **G**. However, in most modern studies, the model is then used to classify areas of geographic space as to whether they present conditions that are in some sense similar to  $\eta(G_+)$  or not – this assessment of similarity establishes whether a site is considered as suitable or not for the species. Hence, in this process, the initial steps are carried out in geographic space, the model is fitted in environmental space, and model outputs are generally visualized in geographic space.

## Species Distribution Modeling or Ecological Niche Modeling?

These seemingly different terms refer to what is in effect the same set of analyses, as can be appreciated from the Introduction and Table 1. In the strictest sense, one could assert that the *target* of model fitting is inevitably some entity in environmental space, which is the realm of niches, and not distributions. In that sense, essentially all modern applications of these techniques should be considered models of ecological niches, which then may be used for a variety of purposes—understanding actual or potential distributional areas, characterizing ecological niches, etc.

Most SDM/ENM studies use something chosen from basically the same set of mathematical algorithms, similar sorts of occurrence data, and the same sets of environmental variables, and both can produce outputs in the form of maps; indeed, papers from the two are published in the same journals. As a consequence, many people fail to perceive that the questions posed in modeling distributions (in a narrow sense) and modeling environments are very different. That is, in SDM and ENM, researchers attempt to calculate very different objects, and a distinction between these two concepts would focus on how geographic areas of interest are identified and characterized. To model a species' distribution, in view of the conceptual framework offered above, one must estimate  $N_p$ , which in turn allows identification of some hypothesis of  $G_A$ . Separately, one must assess **M**, which is often not a simple enterprise (Barve *et al.* 2011). Then, the species' actual (occupied) distributional area is  $G_o = G_A \cap M$ . So, in sum, genuine "species distribution modeling" must include steps of niche estimation *and* steps of assessment of dispersal ability or colonization potential (e.g., van Loon *et al.* 2011). The following section ("When are Species' Distributions Modeled Correctly?") assesses how frequently studies in this field take all of these necessary steps.

Certainly, though, some combinations of these terms and particular applications are not correct. SDM can focus only on the actual distribution – any other use, such as estimating invasive potential or assessing effects of environmental change on species' distributional potential, requires an

explicit estimate of  $N_p$ , and as such must fall under the rubric of ENM. Indeed, of SDM studies, almost half focus on distributional prediction (62 of 126 studies, 49.2%); however, 27.8% focus on climate change projections, 17.5% focus on projections of distributional potential of invasive species, 3.2% make projections of paleodistributional patterns, and 2.4% set out explicitly to characterize the species' ecological niche (!), as is detailed in Table 1. These 50.8% of species distribution models thus use inconsistent and illogical terminology, and should rather be termed ENMs.

## When are Species' Distributions Modeled Correctly?

For those studies that set out to characterize species' geographic distributions (19 ENMs and 62 SDMs), a further question is whether concepts are managed correctly. As detailed above (see "Conceptual Framework and Real-world Implications"), the term 'distribution' has many versions, such that any study purporting to estimate distributions must either (1) restrict model calibration to accessible areas (=M), for which assumptions regarding access must be stated explicitly (Barve *et al.* 2011), thereby estimating  $G_o$  directly; (2) trim modeled hypotheses of  $G_A$  to match hypotheses of accessibility and thereby estimate  $G_o$ ; (3) incorporate dispersal into analyses explicitly (Robinson *et al.* 2011; Boulangeat *et al.* 2012), again estimating  $G_o$  directly; (4) use true absences as a contrast to presences in model calibration and expect that restrictive factors will be captured indirectly as a consequence (Ward *et al.* 2009), or (5) state explicitly that the model outputs are *potential* distributions and not actual distributional area predictions. Hence, we reviewed the methods section of each of the studies that had distributional prediction as a main objective to assess the degree to which they handled the distinction between actual and potential distributions appropriately.

Of the 81 studies, we discarded 16 because they were not empirical in nature and thereby did not confront these issues directly, and we discarded a further 5 studies because access to the full published version was overly difficult, even under the University of Kansas' rather expensive electronic journal access expenditures (such difficult-to-access journals will have to survive without our citations!); hence, we analyzed a total of 60 papers. Of these 60 papers, 32 made no indication of any species-specific considerations as to access to areas or any of the criteria listed above; although some of these analyses were cast on such small spatial extents that access was probably assumed (but not stated explicitly), a large proportion of these studies simply did not state any explicit thought to these considerations. A significant  $2 \times 2$  interaction existed between whether the authors used the SDM *versus* ENM terminology, and whether accounting was made for the difference between potential and actual distributions, with those using SDM tending towards lack of documentation more frequently than expected at random ( $\chi^2 = 10.01$ ,  $df = 1$ ,  $P < 0.002$ ).

## When are Ecological Niches Modeled and Compared Correctly?

Only five studies focused explicitly on estimating and comparing ecological niches out of the sample that we considered (Table 1); of these studies, three used SDM terminology and two used ENM terminology, so no strong tendency was noted. Considering the need for niche estimates and their comparisons to be conditioned on species-specific estimates of M (Barve *et al.* 2011) as discussed above, we assessed each of the five studies as to whether some species-specific area was included in the comparisons.

One paper (Rodder & Engler 2011) discussed metrics of niche overlap and cited key concepts and references, but did not enter into detail about means of comparison and manipulation that they recommended (Warren *et al.* 2008). Of the remaining four papers, three appropriately considered species-specific areas in their niche comparisons (Godsoe 2010b; Nakazato *et al.* 2010; Schulte *et al.* 2012), but one relied on direct comparisons of estimates of niches and distributions, which is quite perilous and likely misleading (Hoisington-Lopez *et al.* 2012).

## Conclusions

We reviewed a representative swath of the recent scientific literature, pondering the degree to which these publications use different terminologies, and the degree to which they hold logical and conceptual inconsistencies. We found considerable cause for concern, such that half of the studies using the terminology "species distribution modeling" nonetheless take advantage of the models for uses related to ecological niches. What is more, about two-thirds of SDM-termed studies and a non-zero number of ENM-termed studies that focus on estimating geographic distributions fail to make appropriate distinctions between actual and potential distributional areas. Finally, one-quarter of a very small sample of studies comparing estimates of ecological niches of species used inappropriate methodologies as well. In sum, an impressive proportion of studies in this emerging field of whole-range species-level ecological biogeography carry logical and conceptual inconsistencies that compromise the rigor of their conclusions.

It is certainly tempting to pass these concerns off as 'just words' (see, e.g., Godsoe 2010a). We argue that this field is compromised by this lack of rigor, however: not only is its credibility in the broader scientific community reduced (Hampe 2004; Sinclair *et al.* 2010), but also many confusions and inappropriate conclusions have been reached for lack of conceptual rigor (Peterson 2011; Soberón & Peterson 2011; Araújo & Peterson 2012). "ENM" should be used only when focus is on estimation of  $N_p$  or  $G_A$  or any potential distribution under changed conditions and circumstances, but care must be taken to distinguish these quantities from their 'existing' or 'realized' manifestations; on the other hand,



“SDM” must include steps to transform areas estimated from potential to actual, so as to reconstruct distributions accurately. If this field is to mature into an important and unique element in the ecological and biogeographic toolkit, greater conceptual rigor will be required.

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